



# The multi-xenobiotic resistance phenotype as a tool to biomonitor the environment

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Organisms use a variety of cellular mechanisms to avoid the effects of toxins. These strategies include de-toxification of putative toxins, sequestration of the toxins or the utilization of transport mechanisms to actually prevent the entry and accumulation of toxins in the cells. These toxin avoidance mechanisms, which presumably evolved in response to natural toxins, can also be used to counter the effects of anthropogenic compounds introduced into the environment by the activities of our modern society. In this article we discuss (1) the use of transport mechanism strategies to protect against toxins and (2) the possible use of these mechanisms as biomarkers indicative of exposure to man-made toxins. We will first review the characteristics of these transport mechanisms, including their biology, genetics and molecular properties and then discuss their use as biomarkers.

**Keywords:** P-glycoprotein, drug resistance, MXR, biomarker, bivalve, mussel.

## The multi-drug resistance phenotype

Multi-drug resistance of tumour cells represents a major problem in cancer treatment, often leading to failure of chemotherapy. Cell lines exhibiting this phenotype, derived by *in vitro* selection with a single cytotoxic drug, show cross-resistance to a variety of structurally and functionally unrelated compounds—hence the term multi-drug resistance for this phenotype (Gottesman and Pastan 1993).

Intensive studies on such cancer cell lines have lead to the realization that there could be transport mechanisms to reduce the accumulation of toxins in eucaryotic cells. Indeed, the main characteristic phenotype of multi-drug resistance (or MDR) is decreased intracellular retention of cytotoxic drugs (Danö 1973, Endicott and Ling 1989). Kinetic studies indicate that this decreased accumulation results from impaired influx and increased efflux outside of the cell, such that the drugs are kept below their toxic level (Gottesman and Pastan 1993).

The most common form of this multi-drug resistance arises from the selection of cells with high amounts of a 170 KDa transport protein referred to as P-glycoprotein or P-gp. This is a membrane-spanning ATP-binding protein that transports chemotherapeutic drugs out of the cells (Juliano and Ling 1976). An absence of definitive substrate specificity is the most characteristic and also the most enigmatic aspect of this transporter. Thus, the transporter effluxes a wide variety of organic compounds, the only common structural feature being moderate hydrophobicity. These substrates include many of the drugs used in cancer

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chemotherapy, many of which are natural products. The list is long and includes such diverse drugs as antibiotics, *Vinca* alkaloids, alkylating agents, anthracyclines and platinum-containing drugs. Furthermore, this mechanism is sensitive to membrane-activity compounds and other chemicals that can specifically block it. These compounds, such as drugs analogues, calcium antagonists or calmodulin inhibitors, are termed modulators or chemosensibilators (see review by Gottesman and Pastan 1993).

Evidence that this multi-drug resistance phenotype arises from over-expression of the P-glycoprotein was provided by transfection experiments with the *MDR1* gene, which showed that over-expression of the MDR-protein alone can confer multidrug resistance on an otherwise drug-sensitive cell line (Gross *et al.* 1986, Ueda *et al.* 1987, Guild *et al.* 1988). It is assumed that this protein acts to reduce accumulation of its substrates in the cell, and the mechanism of reduced accumulation involves the ATP-dependent transport of substrates out of the cell. There are several hypotheses to explain this transport, including the idea that the substrates are pumped from the cytosol or directly from the lipid bilayer (Higgins and Gottesman 1992, Homolya *et al.* 1993) or that the reduced content of drug results from perturbation of the plasma membrane electrochemical potential (Simon *et al.* 1994) which could be related to alterations of the pHi of cells and indirect effects of this pH change on substrate accumulation (Roepe 1995).

### Non-P-gp mediated multi-drug resistance

A number of tumour cell lines display a multiple resistance associated with impaired drug accumulation without P-gp overexpression. In these cases, several proteins of various molecular weight might account for the observed phenotype (Krishnamachary and Center 1993, Scheper *et al.* 1993, Doyle *et al.* 1996). The 110 kDa lung resistance-associated protein (LRP) is associated with an impaired drug accumulation and unfavourable prognosis outcome (Verovski *et al.* 1996, Michieli *et al.* 1997). The multi-drug resistance-associated protein (MRP), is responsible for a drug-resistant phenotype as shown by transfection with the MRP gene (Grant *et al.* 1994, Kruh *et al.* 1994). This 190 kDa MRP protein, is localized in intracellular membranes but predominantly in the plasma membrane and can be differentially glycosylated (Hipfner *et al.* 1994, Zaman *et al.* 1994). The profile of anticancer drugs resistance expelled in the presence of MRP is similar, but not identical, to that of P-gp. It confers resistance to several lipophilic compounds but, unlike P-gp, MRP is able to transport metallic anions including arsenite, arsenate, trivalent and pentavalent antimonials and glutathione conjugates (Cole *et al.* 1994). It is not fully understood how MRP brings about the decreased intracellular drug retention, but it might involve an ATP and glutathione-dependant efflux (Müller M. *et al.* 1994, Zaman *et al.* 1994).

Other mechanisms can be involved in multiple resistance to toxic compounds and these might be of interest to the environmental toxicologist as providing insights as to the spectrum of mechanisms that can be used to resist pollutants. In the 'atypical' MDR phenotype (At-MDR), a broad resistance spectrum is observed without alteration in drug accumulation (Pommier *et al.* 1986, Beck *et al.* 1987, Danks *et al.* 1987). This form of resistance appears to be the result of an alteration in the enzyme DNA topoisomerase II with loss of drug-stimulating DNA cleavage activity (Fernades *et al.* 1990). Some other cases seem to be due to a sum of

different mechanisms including the glutathione S-transferase activity (Zijlstra *et al.* 1987, Townsend and Cowan 1989, Hoban *et al.* 1992).

### The ABC superfamily of transport proteins

Both the MDR and MRP proteins are members of the ATP-binding cassette (ABC) superfamily of transport systems (Hyde *et al.* 1990, Cole *et al.* 1992), also referred to as traffic ATPases (Ames *et al.* 1990). More than 50 members of the family have been identified to date (Higgins 1992). The conserved sequences between all members are short motifs within the ATP-binding domain including the Walker motifs associated with nucleotide binding (Walker *et al.* 1982, Higgins *et al.* 1986) but extend far them, characterizing the ABC domain (Higgins 1992).

Members of the family are widespread among living organisms. In bacteria, transport systems have been found, for example in *Escherichia coli*, in which they can actively import several organic or inorganic nutriments (Zumft *et al.* 1990, Abouhamad *et al.* 1991) and vitamins (Friedrich *et al.* 1986) or export toxins (Felmlee *et al.* 1985, Gilson *et al.* 1990). A large number of other species from different genera, including *Cyanobacteria*, *Mycoplasma*, *Rhizobium*, *Salmonella*, etc. possess a related transporter (reviewed in Higgins 1992). In *Saccharomyces cerevisiae*, the ABC protein is involved in secretion of a pheromone (McGrath and Varshavsky 1989) and several plants have also been shown to possess a member of this family (Ohyama *et al.* 1986, Dudler and Hertig 1992). Among the invertebrates, ABC proteins could be responsible for resistance to several toxic compounds: chloroquine in the malarial parasite *Plasmodium falciparum* (Wilson *et al.* 1989), arsenite (Callahan and Beverley 1991) or drugs in *Leishmania* (Gueiros-Filho *et al.* 1995) and hydrophobic cytotoxic compounds in *Drosophila melanogaster* (Wu *et al.* 1991). In vertebrates, several MDR genes have been found in numerous mammalian species (Ng *et al.* 1989). In humans, two other linked genes have been discovered. One is associated with transport of peptides into the endoplasmic reticulum for class 1 antigen presentation (Müller *et al.* 1994), and another corresponds to the cystic fibrosis gene (CFTR) (Riordan *et al.* 1989).

### MDR-related phenotype in aquatic organisms

Numerous lines of evidence—including isolation of homologous genes, immunological cross-reactivity and activity characteristics—indicate that members of the P-gp (P-glycoprotein) family are present in aquatic organisms.

Winter flounder (*Pleuronectes americanus*) and yellowtail flounder (*Limanda ferrugina*) possess two P-gp genes as indicated by Southern blot analysis (Chan *et al.* 1992) and C-terminal sequences of both genes from *P. americanus* share more than 60% identity to mammalian P-gp genes at the amino acid level. In the sponge *Geodia cydonium*, Northern blot studies using a human cDNA probe revealed the presence of a related 4.2 kb RNA (Kurelec *et al.* 1992). A fragment of 327 bp overlapping the ATP-binding site from *Mytilus edulis* has been found by RT-PCR. Its sequence shows 32% homology with mammalian P-gp gene (Minier 1994). A 650 bp fragment has been sequenced from the marine worm, *Urechis caupo* and shows 60% homology with the human MDR gene (Toomey 1996). Finally, a 130 bp of turbot (*Scophthalmus maximus*) MXR gene shows 80% homology with the human *mdr1* gene (Tutundjian and Minier unpublished results).

An MDR-like protein has also been identified immunologically in a number of aquatic organisms using both polyclonal and monoclonal antibodies. Polyclonal antibodies raised in rabbit against human P-gp reacted with a glycoprotein of 125 kDa in the sponges *G. cydonium* and *Verongia aerophoba* (Kurelec *et al.* 1992) and with a 140 kDa protein from the gastropod *Monodonta turbinata* (Kurelec *et al.* 1995a). The monoclonal antibody C219 has been especially useful since it recognizes a highly conserved region close to the ATP-binding site (Georges *et al.* 1990) and is considered 'a universal probe for the detection of P-glycoprotein' (Endicott and Ling 1989). Using this antibody a 140 kDa protein has been identified in *U. caupo* (Toomey and Epel 1993), two proteins of 220 and 240 kDa in the oyster *Crassostrea gigas* and the blue mussel *M. edulis* (Minier *et al.* 1993), two of 145 and 220–240 kDa in *M. galloprovincialis* (Galgani *et al.* 1995) and one of 170 kDa in *M. californianus* (Cornwall *et al.* 1995). However other reports showed only the 170 kDa protein in *M. edulis* (Minier and Moore 1996). These results could be related to differences in glycosylation or preparation procedures of the protein extracts. Immunohistochemistry has also been used to localize P-gp-related proteins in the shore crab *Carcinus maenas* (Köhler *et al.* 1998a), in the European flounder *Platichthys flesus* (Köhler *et al.* 1998b) and the guppy *Poecilia reticulata* (Hemmer *et al.* 1995).

Photo-affinity analogues have also been used to label the drug transporter. In *U. caupo* oocytes, several proteins could be covalently attached to photoaffinity-labelled forskolin, but only the 140 kDa protein was immunoprecipitated with an antibody to the P-glycoprotein (Toomey and Epel 1993).

MDR-like activity has also been studied in a number of aquatic organisms. The first evidence for an MDR-like activity was drug binding to membrane preparations. Kurelec and coworkers measured the binding of radiolabelled 2-acetylaminofluorene and vincristine to membrane vesicles from molluscs and sponges and found that the extent of labelling was decreased in the absence of ATP, trypsinization of the vesicles or incubation in the presence of a non-radioactive competitive substrate (Kurelec 1992).

Functional evidence for MDR-like activity has come from studies on accumulation of fluorescent substrates such as rhodamine 123, rhodamine B and calcein-AM or with radio-labelled substrates such as vincristine. These studies have found that accumulation of these compounds is enhanced by inhibitors of the MDR-activity (such as verapamil). Examples of such studies include sponge medula (Kurelec 1992, De Flora *et al.* 1995), *Urechis* embryos (Toomey and Epel 1993), *Corbicula* gills (Waldmann *et al.* 1995), mussel gills (Galgani *et al.* 1995) or haemocytes (Minier and Moore 1996) or whole animals (Smital and Kurelec 1997).

## The role of the P-glycoprotein in xenobiotic resistance

As indicated above, the most probable role for the P-glycoprotein transport protein is in protection against xenobiotics. This is obviously the case for the drug-resistant phenotype seen in tumour cells in which the over-expression of the P-gp results in protection from the chemotherapeutic drugs. The P-glycoproteins are also expressed in normal tissues, and their tissue distribution [liver, colon, jejunum, kidney (Thiebaut *et al.* 1987, Sugawara *et al.* 1988)] again suggests a role in protection from diet-derived toxins (or endogene metabolites). In most cases this P-gp is on the apical membranes of cells facing an excretory compartment,

suggesting that the protein activity functions to keep xenobiotics out of the organ or the organism. P-glycoprotein is also expressed at lower levels in the capillary endothelial cells of the brain, testis and placenta (Cordon-Cardo *et al.* 1990) and this localized expression may in part relate to the blood–brain, blood–testis or blood–placenta barriers against xenobiotics.

The most direct evidence indicating a role for these transporters in toxin resistance has come from studies with knockout of genes or inhibition of transport activity. Mice with a knockout for the *mdr-1* gene are normal except for their susceptibility to toxins, such as pesticides (Schinkel *et al.* 1997). P-gp activity is present in mouse embryos, and inhibiting this transport activity enhances sensitivity to the anti-mitotic drug colchicine (Elbling *et al.* 1993). Finally, increasing the dosage of an MDR-related protein in *Caenorhabditis elegans* increased the resistance to colchicine (Broeks *et al.* 1995).

Studies on a variety of aquatic organisms similarly indicate that the natural function of this transporter is in toxin defence. As noted above, the transporter is present in a variety of aquatic organisms and this transport activity indeed functions to protect against xenobiotics. Several studies have shown that this mechanism is induced in response to exposure to toxic compounds (Kurelec 1995, Minier and Moore 1996, Smital and Kurelec 1998). In *M. edulis* haemocytes, rhodamine B retention within the cell was shown to be reduced in relation to the total dose of drug administered and duration of the exposure to the toxic compound vincristine (Minier and Moore 1996). Similarly, incubation of *M. californians* in several different substrates of the P-glycoprotein, such as pentachlorophenol, will induce both protein titre and activity in the mussel gills (Eufemia and Epel 1998). Furthermore, Köhler and coworkers showed that the MXR transporter is expressed in the bile canaliculi of hepatocytes of Flounder (1998b) and in the lysosomal membrane of *C. magneas* hepatocrocraneatic cells (1998a). The later results might indicate the role of MXR in uptake of toxins from the cytoplasm into the lysosomal compartment as seen in mussel haemocytes by confocal laser microscopy (Minier and Moore 1996). Indeed this compartment is of particular importance for toxins tolerance in mussels (Moore and Willows 1998).

This protective mechanism is of particular importance for aquatic organisms as its inhibition might lead to dramatic effects. In embryos of *U. caupo*, for example, cell division is not affected by 0.5  $\mu\text{M}$  vinblastine, but if inhibitors of the transporter are present, cell division ceases (Toomey and Epel 1993). Similarly, genotoxic compounds, as far as they are substrates of the MXR-mechanism, will not greatly affect genome integrity in *C. fluminea* cells at low concentrations. But in the presence of an inhibitor of the MXR transporter, the DNA-damage caused by the genotoxin will be enhanced (Waldmann *et al.* 1995, Kurelec *et al.* 1996b).

This property of making cells resistant to a wide variety of xenobiotics thus provides a multixenobiotic resistance (MXR) which represents a first line of defence for aquatic organisms. The occurrence of inhibitors of the MXR system in the environment raises the concern of the enhancement of toxicity of environmental toxins (Kurelec *et al.* 1998).

### Substrates of the multi-xenobiotic transporter in the aquatic environment

To understand the role of the transporter in toxin protection, one needs to identify the transporter substrates (although the list might be long and always

incomplete). A useful approach to identify substrates (or inhibitors) of the transporter is based on competitive inhibition of the transport activity when two different substrates are present. The easiest way to do this is to measure the effects of a potential substrate on the accumulation of fluorescent compounds such as rhodamine B. Rhodamine B is a substrate for the transporter and the transporter activity will normally exclude the dye and the cell will exhibit little fluorescence. However, if another substrate is present, it will compete with the transporter for the rhodamine, and the concentration of the fluorescent dye will increase in the cell and the cell will be more fluorescent. An increase in fluorescence then indicates that the test compound is a substrate (or inhibitor) of the transport activity. Alternatively, these transport activities and modulation can also be assessed by direct chemical analysis, by accumulation of radioactivity if the compound can be radiolabelled, or by fluorescence measurements with other fluorescent compounds such as rhodamine 123 and calcein-AM.

These approaches were used to identify the pesticide dachtal (dimethyl-2,3,5,6-tetrachloroterephthalate, DCPA) as a substrate of the transporter in mussel gills. When rhodamine and dachtal were incubated together with mussel gill tissue, the fluorescence of the cells increased, indicating that DCPA might be a substrate. This was confirmed by direct chemical analysis of dachtal in experiments in which mussels were incubated in dachtal in the presence or absence of verapamil (the transporter inhibitor). The concentration of pesticide increased two-fold in the presence of the inhibitor (Cornwall *et al.* 1995).

Nevertheless, interpretation of the results of competitive inhibition tests may be complicated by the occurrence of compounds that do not bind to the transporter but inhibit the transport indirectly. For example, staurosporine inhibits the protein kinase C (PKC) which activity is essential for P-gp-mediated transport. Consequently, this inhibitor is not a substrate although it leads to an increased concentration of MXR-substrates (Kurelec 1995). In addition, some results suggest that some substrates can stimulate each other transport by interacting with two cooperative sites within the P-glycoprotein (Shapiro and Ling 1997).

A list of identified substrates (or inhibitors) for the mussel transporter is shown in table 1 which lists substrates that have been identified by (1) competition assays with rhodamine, (2) by direct measurement of accumulation of compounds and enhancement of their uptake by verapamil or (3) by measuring the binding to membrane vesicles prepared from the molluscan tissue. Highly hydrophobic compounds are poor substrates and in *M. californianus*, DDT and Arochlor 1254 did not act as substrates in a rhodamine competition assay but Arochlor (1 ppm) can induce activity and titre of the MXR protein (Eufemia and Epel 1998). In *M. galloprovincialis*, these compounds were weak substrates, showing activity at levels above 3  $\mu\text{M}$ .

Natural waters also contain substrates for the transporter. *In vitro* tests using the rhodamine or calcein-AM competition assay indicated the presence of compounds interacting with the MXR system in pore water from mud flats (Toomey and Epel 1993) and from river and sediments extracts (Kurelec 1995, Kurelec *et al.* 1995b, Smital and Kurelec 1997).

Environmental substrates that are not necessarily toxic can also be present in the environment and can accentuate the effects of more toxic relatives. These less toxic compounds (which could be natural or anthropogenic) could compete with the MXR transport system and saturate the transport mechanism, thus



Table 1. Compounds that interact with the mussel (*M. galloprovincialis* and *M. californianus*) MXR system.

Cytotoxic drugs and modulators	Environmental contaminants
<i>Modulators:</i>	
Verampamil	Dacthal (DCPA)
Forskoline	Chlorbensine
Quinidine	Pentachlorophenol
Trifluoperazine	Sulphate-(N'-(2-chloroallyl)-diethyl-dithiocarbamate (CDEC)
<i>Cytotoxic agents:</i>	
Vinblastine	DDT <sup>a</sup>
Vincristine	DDD <sup>a</sup>
Rhodamine	DDE <sup>a</sup>
2-Acetylaminofluorene	PCB: Arochlor 1254 <sup>a</sup>
Emetine	Hydrocarbons: diesel-2 oil <sup>a</sup>

<sup>a</sup> *M. galloprovincialis* only.

diminishing the effectiveness of the transporter system against more highly toxic substances. The presence of such substrates in natural waters were shown to accentuate the effects of acetoaminofluorene in clam cells (Waldmann *et al.* 1995, Kurelec *et al.* 1996b).

The MXR phenotype as a biomarker

The above results indicate that the MXR system expressed in aquatic organism provides a first line of defence against xenobiotics. It can be induced in response to toxic compounds exposure, indicating that it is a potential biomarker of the health status of the organism and possibly a biomarker to assess the presence of pollutants in the ambient environment.

Measurements of protein titre by Western blots have been made on organisms in the field and these indicate some correlation between amount of MXR protein and organic pollution. In *M. edulis*, specimens from sites highly contaminated with PAH, PCB and DDT in the Seine Bay showed high levels of MXR proteins in gills and mantle tissue, the most contaminated station showing significant higher expression level ( $p < 0.025$ ) (Minier *et al.* 1993). Similarly, MXR protein content assessment in samples from stations in southern England showed significant variation between contaminated and uncontaminated sites (Minier and Moore 1996). In *M. galloprovincialis*, the MXR activity, as assessed by the radiolabelled-vincristine competitive assay, increased with the concentration of pollutants along a pollution gradient (Kurelec *et al.* 1996a). This trend was correlated with the protein expression level with the remarkable exception of the samples from the most contaminated site which showed a marked decrease in transport protein content. In oysters, *Crassostrea virginica*, collected from polluted and unpolluted sites in Charleston Harbor, SC, USA, no significant relationships between MXR proteins and pollutant levels were detected (A. H. Ringwood, personal comm.). And in *Crassostrea gigas*, no relationships between MXR proteins and pollution level were found along the French south-west coast (Minier *et al.* 1993, Minier 1994). These later results might indicate differences between species or interaction of other factors, as discussed below.

If MXR protein measurements can be validated as a marker, the available immunological reagents, such as the C219 antibody, will prove simple to use. This antibody, directed against mammalian P-gp, apparently works against the P-gp present in a wide variety of organisms. Western blots are tedious but dot blot or slot blot measurements could be carried out on a routine basis if Western blot analysis indicates that there is no cross-reaction with other proteins for the particular test tissue (Galgani *et al.* 1995, Minier unpublished results).

Activity tests should also be done in concert with the above immunological assays but these are more difficult to perform. Simple methods are needed for this purpose. Accordingly, Kurelec and coworkers developed a rhodamine B-efflux method using whole animals. Mussels are exposed to rhodamine-containing water and then transferred to clean water. Rate of efflux can then be assessed in a non-destructive manner (Smital and Kurelec 1997). Another approach is to use calcein-AM as a substrate and to measure its accumulation in hemocytes (Minier and Moore 1998).

Interpreting *in situ* measurements of the MXR phenotype might not be simple, however. One concern is that the MXR phenotype can be induced in response to natural factors that are not necessarily indicative of pollution. For example, the mammalian MDR system is affected by heat shock (Chin *et al.* 1990) and modulated by several hormones (Qian and Beck 1990). In a similar manner, heat shock can induce MXR protein expression in mussels (Eufemia and Epel unpublished) and results of a 1 year study of mussel MXR protein and heat shock protein (HSP70) expression showed that they vary similarly during the year with a peak in August and are correlated with water temperature (Minier *et al.* 1999).

A related and potentially more serious problem is separating out the effects of natural toxins from pollutants. For example, bacteria and plants produce substrates for the MXR transporter, and seasonal variations in levels of these natural products could induce concomitant increases in MXR titre and activity (Toomey *et al.* 1996). Indeed, macroalgal fragments and phytoplankton present in Monterey Bay produce transport substrates (Eufemia and Epel unpublished) and, for example, extracts from the marine algae, *Caulerpa taxifolia*, can reverse the multi-xenobiotic resistance (Smital *et al.* 1996). Variations in content of these substrates could induce the transporter of pollutant concentrations. Indeed, the observed and seasonal variations in the titre activity of the transporter could be related to variations in natural substrates as well as to variations in temperature.

A related problem comes from the finding that there can also be seasonal variation in inducibility of the MXR protein in the gill tissue of *M. californianus* in response to substrates (Eufemia and Epel, unpublished). This inability to induce the MXR response has been particularly seen when the titre is already high, as in the summer months. This could result from a saturation response, in which the amount of transporter protein reaches a maximum such that no further induction can occur. This variation in inducibility has been especially apparent this past year (1997/98), which was during an intense El Nino cycle. The variations could be related to higher water temperature, as noted above. Other factors that could increase the titre could be sporadic algal blooms. An opposite factor that could affect the MXR response could be nutrient availability, with poor or rich nutrient conditions affecting the physiology and response of the mussels. There could also be effects related to the reproductive status of the animals, but these have not yet been studied.



Interaction of natural parameters are common to many biological responses and, in the case of MXR, further work should focus on the possible way(s) to correct or adjust for these sources of variation. Two possible ways of interest can be mentioned in order to circumvent the above problems. One is to screen other organisms to find ones that are not so subject to seasonal variation. In this regard, mussels are typically harvested from intertidal areas, and these might be more susceptible to seasonal stresses than benthic animals. It could be worth surveying a variety of organisms from different habitats to see if there are bivalves (or other test organisms) that normally have a low level of MXR protein but which is inducible in response to xenobiotics.

*In vitro* tests using animals or cells grown in the laboratory could provide a second way to avoid problems arising from changes in season or reproductive status. One could use these animals or cell lines to measure induction in response to exposure to test water samples. Cell lines of marine invertebrates would be ideal for these types of studies, but good sources of such cells are not available, and this is a subject for future research. Having such lines would considerably simplify testing of waters for their ability to induce the MXR protein. They could also be used to determine the cellular and molecular mechanisms of sensing pollutants and the cell response to these compounds.

A final unanswered question is whether other transport mechanisms besides the MXR strategy is used by aquatic organisms to avoid the effects of toxins. In this paper we mentioned the MRP transporter, which is one example of a newly-discovered transport protein in mammals, that might be of general importance to all organisms. There may be many members of such transport families, and insights into these mechanisms could provide alternative candidates for useful biomarkers.

## Conclusions

The multi-xenobiotic resistance system is an important defence mechanism against environmental toxins which has been highly conserved during evolution. This phenotype can be induced by exposure to toxic compounds and significant different expression levels can be assessed between populations living in more or less polluted sites. Accordingly, this phenotype has the potentials of a biomarker which might be useful for environmental diagnosis. However, as for many biological responses, natural factors can interfere with the MXR phenotype. Heat shock and compounds of natural origin can modulate both activity and protein expression and future work should address the important question of finding ways to correct for these sources of variation.

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